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Application of Alumina-coated Carbon Nanotubes in Removal of Tetracycline from Aqueous Solution

Davoud Balarak^{1*}, Yousef Mahdavi² and Ferdos Kord Mostafapour¹

¹Department of Environmental Health, Health Promotion Research Center, Zahedan University of Medical Sciences, Zahedan, Iran. ²Department of Environmental Health Engineering, Student Research Committee, Mazandaran University of Medical Sciences, Sari, Iran.

Authors' contributions

This work was carried out in collaboration between all authors. All authors read and approved the final manuscript.

Article Information

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ABSTRACT

The alumina-coated multi walled carbon nanotubes (AI_2O_3 -coated MWCNTs) was used for removal of Tetracycline (TC) from aqueous solution in batch systems. The adsorption of TC occurred by studying the effects of adsorbent amount, TC concentration, contact time and temperature. The results suggested the adsorption capacity increased with increasing temperature. The equilibrium adsorption data were analyzed using three widely applied isotherms: Langmuir, Freundlich and Temkin. The results revealed that Langmuir isotherm fit the experimental results well. Kinetic analyses were conducted using pseudo-first and second-order models and the intraparticle diffusion model. The regression results showed that the adsorption kinetics were more accurately represented by pseudo-second-order model.

Keywords: Tetracycline; alumina-coated multi-wall carbon nanotube; adsorption; kinetics; isotherms.

*Corresponding author: E-mail: dbalarak2@gmail.com;

1. INTRODUCTION

In the last decades, the presence of emerging contaminants in different types of water appeared as a new environmental threat which needs to be faced by several governments around the word [1,2]. Pharmaceutical substances and personal care products are an emerging class of aquatic contaminants that have been increasingly detected in ground and surface water [3,4]. Generally, pharmaceuticals reach waterways through the discharge of wastewaters and effluents on environment, which often are not properly treated [5]. On the other hand, it was proved that when pharmaceuticals reach sewage wastewater treatment plants (SWTP), they are not completely removed [6,7]. The tetracycline (TC) group is one of the most commonly used antibiotics worldwide, which finds wide applications in human therapy and the livestock industry, and is difficult to be metabolized [8,9]. A large extent (up to 72%) of TC was excreted with unmetabolized form from human beings and animals [10,11]. Removing antibiotics is difficult and requires expensive process. Physical and physico-chemical techniques have proved their efficiency for this purpose [10]. Among the Physical techniques adsorption. flocculation, electro-flocculation, reverse osmosis, ultrafiltration, coagulation have been applied to remove recalcitrant pollutants [12,13]. Among the techniques, adsorption of TC to different sorbents is believed the easiest and cheapest one [14]. Inorganic (aluminum oxide [15], Fe-Mn binary oxide [16], zeolite beta [17], goethite [18], montmorillonite [19,20], cinnamon soil [21], kaolinite [22], illite [23] and swelling clay minerals [24], biological (algal ponds [25], nitrifying granules [26], bio-char [27], activated sludge system [28], synthetic (chitosan [29], and composite sorbents (MnFe₂O₄/ activated carbon magnetic composite [30], titania-silica composed materials [31] are widely used for the adsorptive removal of TCs. Nanotechnology has been successfully applied in the field of pollution abatement to remove of toxic contaminants, remediate of groundwater, etc [32,33]. The their high specific surface area, environmental benignity and capabilities as catalytic degradation of contaminants through changing reaction mechanism [34,35]. Due to their large specific surface area, small size, and hollow and lavered structures, Carbon Nano Tube (CNT) have been proven to possess great potential as superior adsorbents for removing many kinds of organic and inorganic contaminants [36,37]. The uses of CNT as support of metallic oxides have

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still been reported. The advantages of this chemistry are that materials with higher sorption capacity, larger surface area and supported metallic oxides with better orientation degree with regard to net metallic oxides can be obtained [15,38]. This work aims to investigate the adsorption behavior of alumina supported on carbon nanotube for the treatment of TC aqueous solutions. For this purpose MWCNT/Al₂O₃ composites were synthesized and characterized. The effect of contact time, dosage of adsorbent and initial concentration of TC was studied.

2. MATERIALS AND METHODS

2.1 Chemicals

The tetracycline hydrochloride (Molecular weight: 480.9, Molecular formula: $C_{22}H_{24}N_2O_8$ ·HCl) was purchased from Sigma-Aldrich, USA. The chemical structure of Tetracycline is presented in Fig. 1. The distilled water was used to prepare the stock solution of tetracycline. Other chemical used in this study were prepared from Merck, Germany.



Fig. 1. The chemical structure of tetracycline hydrochloride

2.2 Adsorbent Preparation

The multi-wall carbon nanotubes (MWCNTs) used in this study were of more than 98% purity and provided from Research Institute of Petroleum Industry (RIPI), Tehran, Iran). Further purification of the MWCNTs was accomplished by stirring them in concentrated nitric acid at 70°C for 12 h, followed by filtering and washing with distilled water, and then drying at 110°C for 6 h. Then, the MWCNTs were oxidized by refluxing with 50% nitric acid at 120°C for 12 h under stirring conditions. The product was then filtered and rinsed with doubly distilled water and dried overnight in the oven. The preparation of

Al₂O₃/MWCNT composites was accomplished as follows: All glassware was cleaned by agua regia freshly prepared prior to use. Typically, some amounts of MWCNTs were dispersed into deionized water and magnetically agitated 6 h at which acceptable level of dispersion was observed. A specific amount of aluminum nitrate was properly dissolved in deionized water. Carefully, the aluminum nitrate solution was drop wise added into dispersed MWCNTs. Between consecutive drops, there should be an appropriate time for the aluminum to reach, appropriately disperse and engage to the MWCNTs' surface. After that, the suspension was dried at 110°C. The obtained material was heated up to 400℃ for 90 min, where the pyrolysis process resulted in alumina formation supported onto the MWCNTs' surface [39,40].

The surface area of the samples were determined by using Micromeritics ASAP 2020 volumetric adsorption analyzer. The BET surface area was measured from the adsorption isotherm using Brunauer–Emmett–Teller equation. The surface morphology of the MWCNTs and Al_2O_3 -coated MWCNTs were examined using a environmental scanning electron microscopy (ESEM) instrument (Philips XL30 ESEM-TMP) and also a X-ray diffraction spectrometer (Philips-X'Pert Pro MPD) with the conditions: generator voltage of 40 kV, tube current of 40 mA, scan step size of 0.02°, and the scan range of 5–90°.

2.3 Adsorption Experiment

Adsorption experiments were conducted by contact time, adsorbent dose. varying temperature, initial TC concentration at pH fixed 7 under the aspects of thermodynamic study, adsorption isotherms and adsorption kinetics. The experiments were carried out in 200 mL Erlenmeyer flasks and the total volume of the reaction mixture was kept at 100 mL. The pH of solution was maintained at a desired value byadding 0.1 M NaOH or HCl. The flasks were shaken for the requiredtime period in a water bath shaker. The kinetic studies was carried out by agitating 200 mL flasks containing 1 g/L of adsorbent and 100 mL TC solutions of different concentrations i.e. from 10 to 100 mg/L, in water bath shaker. The mixture was agitated at 180 rpm at 25°C. The contact time was varied from 5 to 120 min. The initial and equilibrium concentration of TC in solution was determined by HPLC. In the HPLC analysis and a SPD-10A UV–Vis detector at maximum absorption wavelength of 365 nm which was detemined using a Shimadzu UV-1700 spectrophotometer to scan from 200 to 800 nm. A 30:70 (v/v) acetonitrile and 0.01 M aqueous oxalic acid mixture was used as mobile phase at room temperature with a constant flow rate of 1.0 mL min⁻¹. The injection volume was 10 ml. The retention time of TC hydrochloride was 4 min. All of adsorption experiments in this study were performed in triplicate and the results were given as average.

The adsorption capacity of the Al_2O_3 -coated MWCNTs at the equilibrium state was calculated as follows [41,42]:

$$q_e = \frac{(C_0 - C_e)V}{M}$$

where q_e (mg/g) is the TC adsorption capacity, V (L) is the volume of solution, M (g) is the weight of adsorbent, C_0 and C_e (mg/L) denote the initial and equilibrium concentrations of TC, respectively. All experiments were conducted in duplicates and the average values were adopted for the analysis.

3. RESULTS AND DISCUSSION

The MWCNT surface presents polar groups, such as hydroxyl or carboxyl groups which are able to interact, by hydrogen bonding with the Al_2O_3 surface. The presence of these groups onto MWCNT surface is consequence of acid treatment employed during synthesis process to dissolve the catalysts [40,43]. The Al_2O_3 wrapping layer surrounds the MWCNT surface, in which the oxide layer stays fixed. However, some agglomerate Al_2O_3 particles were also observed in the material, as can be seen in Fig. 2. Also specific surface area of the adsorbent used in this study was 441.2 m²/g.

Figs. 3a and 3b shows the XRD patterns of MWCNTs and Al_2O_3 -coated MWCNTs. It was observed that the two peaks corresponding to the structure of MWCNTs (Fig. 3a) also exist in the XRD pattern of the Al_2O_3 -coated MWCNTs (Fig. 3b). The broad peak of MWCNTs in XRD was seen narrow in the XRD of the composite because MWCNT in the composite was coated with alumina. The observed XRD of the synthesized Al_2O_3 -coated MWCNTs was in convenience with that reported for the synthesis of alumina carbon nanotube composite using thermal pressure method.



Fig. 2. SEM photograph of Al₂O₃-coated MWCNTs

3.1 Effect of Adsorption Contact Time and Initial TC Concentration

Effect of contact time for the removal of TC by the Al₂O₃-coated MWCNTs at Co=10 to 100 mg/L for the adsorbent dose of 1 g/L showed rapid adsorption of TC in the first 30 min and, thereafter, the adsorption rate decreased gradually and the adsorption reached equilibrium in about 60 min as shown in Fig. 4. Initially, a large number of vacant surface sites are available for adsorption. The adsorption rate is very fast thus rapidly increases the amount of adsorbates accumulated on the Al₂O₃-coated MWCNTs surface mainly within the first 30 min of adsorption. As a result, the remaining vacant surface sites are difficult to be occupied due to formation of repulsive forces between the TC molecules on the solid surface and in the bulk phase [44,45]. Also Fig. 4 describes the effect of TC initial concentrations on the removal percentage by Al₂O₃-coated MWCNTs. The TC removal percentage via Al₂O₃-coated MWCNTs was found to decrease with increase in initial concentrations. This may be due to the vacant sites are filled up and no further adsorption occurs due to saturation of vacant sites of adsorbent [46,47].

3.2 Effect of Adsorbent Dosage

The adsorption percent at various doses of Al_2O_3 -coated MWCNTs from 0.01 to 2 g/L is shown in Fig. 5. Increasing the adsorbent dose to 1 g/L increase the adsorption percent of TC, which is due to the increasing in adsorption sites of adsorbent material resulting from increasing of

surface area of adsorbent [48]. However, further increase of adsorbent dosage does not afford exhaustive adsorption of TC. This may be due to overlapping of adsorption sites as a result of overcrowding of adsorbent particles [49].

3.3 Adsorption Isotherms Study

Equilibrium studies that give the capacity of the adsorbent and adsorbate are described by adsorption isotherms, which is usually the ratio between the quantity adsorbed and that remained in solution at equilibrium at different temperature. Freundlich, Langmuir and Temkin isotherms are the earliest and simplest known relationships describing the adsorption equation.

3.4 Langmuir Isotherm

The Langmuir isotherm is derived on the assumption of mono-layer adsorption on a homogenous surface. It is expressed by [50,51]:

$$\frac{Ce}{q_e} = \frac{1}{q_m K_L} + \frac{C_e}{q_m}$$

where C_e is the equilibrium concentration (mg/L), q_e is the amount of TC adsorbed (mg/g), q_m is q_e for complete monolayer adsorption capacity (mg/g), and K_L is the equilibrium adsorption constant (L/mg).

The essential characteristic of a Langmuir isotherm can be expressed in terms of a dimensionless separation factor, R_{L} [52]:

$$R_{L} = \frac{1}{1 + K_{L}C_{O}}$$

where K_L is the Langmuir constant and C_0 is the lowest initial TC concentration (mg/L).

The Freundlich isotherm is suitable for non-ideal adsorption on heterogeneous surfaces. The heterogeneity is caused by the presence of different functional groups on the surface, and several adsorbent–adsorbate interactions. The Freundlich isotherm is expressed by the following empirical equation [53,54]:

$$\text{Log } q_e = \frac{1}{n} \log Ce + \log K_F$$

Where q_e is the equilibrium TC concentration on the adsorbent (mg g⁻¹); C_e, the equilibrium TC concentration in solution (mg L⁻¹); and K_F is the Freundlich constant.

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Fig. 3. X-ray diffraction of uncoated MWCNTs and Al₂O₃-coated MWCNTs



Fig. 4. Effect of contact time and initial TC concentration (pH =7, Adsorbent dosage 0.1 g/L and temperature = 25±3°C)



Fig. 5. Effect of adsorbent dosage on TC adsorption onto Al_2O_3 -coated MWCNTs ($C_0 = 100 \text{ mg/L}$, time = 60 min, pH = 7, temperature = $25\pm3^{\circ}C$)

The Temkin equation suggests a linear decrease of sorption energy as the degree of completion of the sorptional centers of an adsorbent is increased. The heat of adsorption of all the molecules in the layer would decrease linearly with coverage due to adsorbent-adsorbate interactions. The adsorption is characterized by a uniform distribution of binding energies. The Temkin isotherm has been generally applied in the following form [55,56]:

$$q_e = \frac{RT}{b} \ln (AC_e)$$

And it can be linearized as [57]:

$$q_e = B \ln A + B \ln C_e$$

where B = RT/b, b is the Temkin constant related to heat of sorption (J/mol), A is the Temkin isotherm constant (L/g), R is the gas constant (8.314 J/(mol K)) and T is the absolute temperature (K).

The equilibrium isotherms was studied by varying the initial concentration of TC at temperure 273, 298 and 323 K under the conditions of pH 7.0, contact time of 60 min and adsorbent dose 0.1 g/L. The fitting results, i.e. isotherm parameters and the coefficient of determination, R², were shown in Table 1. The value of R_L for adsorption of TC indicated that the adsorption behavior of Al₂O₃-coated MWCNTs was favorable for the TC (R_L<1). According to a linear regression method, the Freundlich and Temkin isotherms were poorly suitability to the adsorption of AI_2O_3 coated MWCNTs in comparison to Langmuir isotherm. The validity of the Langmuir model suggested the TC uptake was due to monolayer coverage of solute particles onto the surface of the activated carbon and adsorption of each molecule has equal activation energy. All three models suggest that increasing temperature increased adsorption capacity, revealing that the adsorption is endothermic. The same behavior was observed by Balarak et al. [7].

3.5 Adsorption Kinetics

In order to examine the controlling mechanism of adsorption processes such as mass transfer and chemical reaction, pseudo-first-order, pseudosecond-order and intraparticle diffusion kinetic equations were used to test the experimental data. The pseudo-first-order kinetic model has commonly been applied in the following form [58]:

Log
$$(q_e - q_t) = \log q_e - \frac{K_1}{2.303}t$$

Where q_t is the adsorption capacity at time t (mg/g) and k_1 (min⁻¹) is the rate constant of the pseudo- first order adsorption, was applied to the present study of TC adsorption. The rate constant, k_1 and correlation coefficients of the TC under different concentrations were calculated from the linear plots of $ln(q_e -q_t)$ versus t (Fig. 6a) and listed in Table 2. The correlation coefficients for the pseudo-first-order kinetic model are low. Moreover, a large difference of

equilibrium adsorption capacity (q_e) between the experiment and calculation was observed, indicating a poor pseudo-first-order fit to the experimental data.

The pseudo-second-order kinetic model is expressed as [59]:

$$\frac{\mathrm{t}}{\mathrm{q}_{\mathrm{t}}} = \frac{\mathrm{1}}{h} + \frac{\mathrm{t}}{\mathrm{q}\mathrm{e}}$$

where $h = kq_e^2$ (mg g⁻¹ min⁻¹) can be regarded as the initial adsorption rate as $t \rightarrow 0$ and k is the rate constant of pseudo-second- order adsorption (g $mg^{-1}\ min^{-1}).$ The plot t/qt versus t (Fig. 6b) should give a straight line if pseudosecond-order kinetics is applicable and q_e , k and h can be determined from the slope and intercept of the plot, respectively. At all studied initial TC concentrations, the straight lines with extremely high correlation coefficients (> 0.997) were obtained. In addition, the calculated q_e values also agree with the experimental data in the case of pseudo- second-order kinetics. These suggest that the adsorption data are well represented by pseudo-second-order kinetics and this supports the assumption that the rate-limiting step of TC onto Al₂O₃-coated MWCNTs may be chemical sorption or chemisorption. From Table 2, the values of the rate constant k decrease with increasing initial TC concentration for the Al₂O₃coated MWCNTs. The reason for this behavior can be attributed to the lower competition for the sorption surface sites at lower concentration. At higher concentrations, the competition for the surface active sites will be high and consequently lower sorption rates are obtained. Similar phenomena have been observed in the adsorption of amoxicillin on Single Walled Carbon Nanotubes [60,61].

In order to gain insight into the mechanisms and rate controlling steps affecting the kinetics of adsorption, the kinetic experimental results were fitted to the Weber's intraparticle diffusion. The kinetic results were analyzed by the intraparticle model to elucidate the diffusion mechanism, which model is expressed as [62,63]:

$$q_t = k t^{1/2} + C$$

Where C is the intercept and k is the intraparticle diffusion rate constant, (mg/gmin^{1/2}), which can be evaluated from the slope of the linear plot of q_t versus $t^{1/2}$ as shown in Fig. 6c. The intercept of the plot reflects the boundary layer effect. The larger the intercept, the greater the contribution of the surface sorption in the rate controlling

step. The calculated intraparticle diffusion coefficient k values are listed in Table 2. The lf the regression of q_t versus $t^{1/2}$ is linear and passes through the origin, then intraparticle diffusion is the sole rate-limiting

step. However, the linear plots at each concentration did not pass through the origin. This indicates that the intraparticle diffusion was not only a rate controlling step.



Fig. 6. Biosorption kinetics (a): pseudo first-order (b): pseudo-second-order (C): intraparticle diffusion

Langmuir model						Freundlich model			Temkin model		
Temp (K)	q e exp	q m	ΚL	R∟	R ²	n	K _F	R ²	В	Α	R ²
273	851.2	833.4	0.012	0. 45	0.998	2.81	59.36	0.842	47.69	12.84	0.799
298	874.3	855.6	0.031	0.59	0.997	3.17	72.11	0.858	53.46	9.424	0.808
323	895.7	876.8	0.056	0.74	0.999	3.64	84.48	0.817	61.87	7.436	0.824

Table 1. The adsorption isotherms constants for the removal TC

Table 2. Kinetic parameters	for	TC adsorption	on onto A	Al ₂ O ₃ -coated	MWCNTs
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	Pseudo second-order model				Pseud	lo first-orc	ler model	Intraparticle diffusion		
C。	q _e ,exp (mg/g)	k ₂	R ²	q e cal	K 1	R ²	q e cal	К	С	R ²
10	99.4	0.0041	0.997	94.46	0.094	0.841	56.24	0.061	0.805	0.736
25	247.75	0.0059	0.998	249.8	0.042	0.823	214.7	0.122	1.245	0.759
50	485.5	0.0074	0.998	474.2	0.035	0.796	402.8	0.179	1.517	0.784
100	901.2	0.0088	0.997	889.5	0.019	0.803	745.2	0.252	1.985	0.802

4. CONCLUSION

The batch removal of TC from aqueous solution by using Al_2O_3 -coated MWCNTs was studied. The effect of temperure, initial concentration of TC, Al_2O_3 -coated MWCNTs content and contact time on the removal of TC was also investigated. The Langmuir adsorption isotherms are found to provide the best fit to the experimental data with a maximum adsorption capacity of 855.6 mg/g at room temperure. The adsorption kinetics can be determined by pseudo-second-order kinetics. The results of the present investigation indicate that the Al_2O_3 -coated MWCNTs has potential for use in removing TC antibiotic from aqueous solutions.

CONSENT

It is not applicable.

ETHICAL APPROVAL

It is not applicable.

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COMPETING INTERESTS

Authors have declared that no competing interests exist.

REFERENCES

1. Zazouli MA, Mahdavi Y, Bazrafshan E, Balarak D. Phytodegradation potential of bisphenolA from aqueous solution by *Azolla filiculoides*. Journal of Environmental Health Science & Engineering. 2014;12(66):1-5.

- Diyanati RA, Yazdani J, Balarak D. Effect of sorbitol on phenol removal rate by lemna minor. Mazandaran University of Medical Science. 2013;22(87):58-64.
- Ghauch A, Tuqan A, Assi HA. Elimination of amoxicillin and ampicillin by micro scale and nano scale iron particles. Environ Pollut. 2009;157:1626–1635.
- Zhang W, He G, Gao P, Chen G. Development and characterization of composite nanofiltration membranes and their application in concentration of antibiotics. Sep Purif Technol. 2003;30:27– 35.
- Gulkowsk A, Leung HW, So MK, Taniyasu S, Yamashita N. Removal of antibiotics from wastewater by sewage treatment facilities in Hong Kong and Shenzhen, China. Water Research. 2008;42:395-403.

 Balarak D, Mahdavi Y, Azadi NA, Sadeghi SH. Isotherms and thermodynamic study on the biosorption of amoxicillin using canola. International Journal of Analytical, Pharmaceutical and Biomedical Sciences. 2016;5(3):8-14.

 Balarak D, Kord Mostafapour F, Joghataei A. Experimental and kinetic studies on penicillin g adsorption by lemna minor. British Journal of Pharmaceutical Research. 2016;9(5):1-10.

 Yahiaoui I, Aissani-Benissad F, Fourcade F, Amrane A. Removal of tetracycline hydrochloride from water based on direct anodic oxidation (Pb/PbO2 electrode) coupled to activated sludge culture. Chemical Engineering J. 2013;221:418-425.

- Ocampo-Pérez R, Rivera-Utrilla J, Gómez-Pacheco C, Sánchez-Polo M, López-Peñalver JJ. Kinetic study of tetracycline adsorption on sludge-derived adsorbents in aqueous phase. Chem Eng J. 2012;213:88–96.
- Parolo ME, Savini MC, Vallés JM, Baschini MT, Avena MJ. Tetracycline adsorption on montmorillonite: pH and ionic strength effects. Appl. Clay Sci. 2008;40:179–186.
- 11. Brinzila CI, Pacheco MJ, Ciríaco L, Ciobanu RC, Lopes A. Electro degradation of tetracycline on BDD anode. Chem. Eng. J. 2011;209:54–61.
- Martins AC, Pezoti O, Cazetta AL, Bedin KC, Yamazaki DAS, Bandoch GFG. Removal of tetracycline by NaOH-activated carbon produced from macadamia nut shells: Kinetic and equilibrium studies. Chemical Engineering Journal. 2015;260;291-299.
- 13. Choi KJ, Kima SG, Kimb SH. Removal of antibiotics by coagulation and granular activated carbon filtration. Journal of Hazardous Materials. 2008;151:38–43.
- Malakootian M, Balarak D, Mahdavi Y, Sadeghi SH, Amirmahani N. Removal of antibiotics from wastewater by *Azolla filiculoides*: Kinetic and equilibrium studies. IJAPBS. 2015;4(7):105-113.
- Chen WR, Huang CH. Adsorption and transformation of tetracycline antibiotics with aluminum oxide. Chemosphere. 2010; 79:779–785.
- Liu H, Yang Y, Kang J, Fan M, Qu J, Removal of tetracycline from water by Fe– Mn binary oxide. J. Environ. Sci. 2012; 24:242–247.
- Kang J, Liu H, Zheng YM, Qu J, Paul Chen J. Application of nuclear magnetic resonance spectroscopy, fourier transform infrared spectroscopy, UV–Visible spectroscopy and kinetic modeling for elucidation of adsorption chemistry in uptake of tetracycline by zeolite beta. J. Colloid Interface Sci. 2011;354:261–267.
- Zhao Y, Geng J, Wang X, Gu X, Gao S. Adsorption of tetracycline onto goethite in the presence of metal cations and humic substances. J. Colloid Interface Sci. 2011; 361:247–251.
- Eugenia Parolo M, Avena MJ, Pettinari GR, Baschini MT. Influence of Ca2+ on tetracycline adsorption on montmorillonite. J. Colloid Interface Sci. 2012;368:420–426.

- Zhao Y, Gu X, Gao S, Geng J, Wang X. Adsorption of tetracycline (TC) onto montmorillonite: Cations and humic acid effects. Geoderma. 2012;183;12–18.
- 21. Wan Y, Bao Y, Zhou Q. Simultaneous adsorption and desorption of cadmium and tetracycline on cinnamon soil. Chemosphere. 2010;80:807–812.
- 22. Li Z, Schulz L, Ackley C, Fenske N. Adsorption of tetracycline on kaolinite with pH-dependent surface charges. J. Colloid Interface Sci. 2010;351:254–260.
- Chang PH, Li Z, Jean J, Jiang W, Wang CJ, Lin KH. Adsorption of tetracycline on 2:1 layered non-swelling clay mineral illite. Appl. Clay Sci. 2012;67–68:158–163.
- 24. Chang PH, Li Z, Jiang WT, Jean JS. Adsorption and intercalation of tetracycline by swelling clay minerals. Appl. Clay Sci. 2009;46:27–36.
- 25. de Godosa I, Munozb R, Guieyssea B. Tetracycline removal during wastewater treatment in high-rate algal ponds. J. Hazard. Mater. 2012;229–230:446-449.
- Shi YJ, Wang XH, Qi Z,Diao MH, Gao MM, Xing SF, Wang SG, Zhao XC. Sorption and biodegradation of tetracycline by nitrifying granules and the toxicity of tetracycline on granules. J. Hazard. Mater. 2011;191:103–109.
- Liu P, Liu W, Jiang H, Chen J, Li W, Yu H. Modification of bio-char derived from fast pyrolysis of biomass and its application in removal of tetracycline from aqueous solution. Bioresour. Technol. 2012;121:235–240.
- Prado N, Ochoa J, Amrane A. Biodegradation and biosorption of tetracycline and tylosin antibiotics in activated sludge system. Process Biochem. 2009;44:1302–1306.
- 29. Kang J, Liu H, Zheng Y, Qua J, Paul Chen J. Systematic study of synergistic and antagonistic effects on adsorption of tetracycline and copper onto a chitosan. J. Colloid Interface Sci. 2010;344:117–125.
- Shao L, Ren Z, Zhang G, Chen L. Facile synthesis, characterization of a MnFe2O4/ activated carbon magnetic composite and its effectiveness in tetracycline removal. Mater Chem Phys. 2012;135:16–24.
- Brigante M, Schul PC. Remotion of the antibiotic tetracycline by titania and titania– silica composed materials. J. Hazard. Mater. 2011;192:1597-1608.
- 32. Upadhyayula VKK, Deng S, Mitchell MC, Smith GF. Application of carbon nanotube

technology for removal of contaminants in drinking water: A review. Science of the Total Environment. 2009;408:1–13.

- Yang K, Wu W, Jing Q, Zhu L. Aqueous adsorption of aniline, phenol and their substitutes by multi-walled carbon nanotubes. Environ Sci Technol. 2008;42:7931-6.
- 34. Yao Y, He B, Xu F, Chen X. Equilibrium and kinetic studies of methyl orange adsorption on multiwalled carbon nanotubes. Chemical Engineering Journal. 2011;170:82–89.
- 35. Ji LG, Chen W, Duan L, Zhu D. Mechanisms for strong adsorption of tetracycline to carbon nanotubes: A comparative study using activated carbon and graphite as adsorbents. Environmental Sciences Technology. 2009;43:2322– 2327.
- 36. Balarak D, Mahdavi Y, Bazrafshan E, Mahvi AH, Esfandyari Y. Adsorption of fluoride from aqueous solutions by carbon nanotubes: Determination of equilibrium, kinetic and thermodynamic parameters. Flouride. 2016;49(1):35-42.
- Yao Y, Xu F, Chen M, Xu Z, Zhu Z. Adsorption behavior of methylene blue on carbon nanotubes. Bioresource Technology. 2010;101:3040–3046.
- Gao Y, Li Y, Zhang L, Huang H, Hu J, Shah SM, Su X. Adsorption and removal of tetracycline antibiotics from aqueous solution by graphene oxide. J. Colloid. Interface Sci. 2012;368:540–546.
- 39. Renata S. Amaisa, Julia na S. Ribeiro a, Mariana G. Segatelli b, Inez VP. Yoshida b, Pedro O. Luccas a, C'esar RT. Assessment of nanocomposite alumina supported on multi-wall carbon nanotubes as sorbent for on-line nickel preconcentration in water samples. Separation and Purification Technology. 2007;58:122–128.
- 40. Gupta VK, Agarwal S, Saleh TA. Synthesis and characterization of alumina-coated carbon nanotubes and their application for lead removal. Journal of Hazardous Materials. 2011;185:17-23.
- 41. Zazouli MA, Yazdani J, Balarak D, Ebrahimi M, Mahdavi Y. Removal acid blue 113 from aqueous solution by canola. Journal of Mazandaran University Medical Science. 2013;23(2):73-81.
- 42. Zazouli MA, Balarak D, Karimnezhad F, Khosravi F. Removal of fluoride from aqueous solution by using of adsorption

onto modified Lemna minor: Adsorption isotherm and kinetics study. Journal of Mazandaran University Medical Sciences 2014;23(109):208-17.

- 43. He CN,Tian F, Liu SJ. A carbon nanotube/ alumina network structure for fabricating alumina matrix composites. Journal of Alloys and Compounds. 2009;478:816– 819.
- 44. Aksu Z, Tunc O. Application of biosorption for penicillin G removal: Comparison with activated carbon. Process Biochemistry. 2005;40(2):831-47.
- 45. Rivera-Utrilla J, Prados-Joya G, Sánchez-Polo M. Removal of nitroimidazole antibiotics from aqueous solution by adsorption/bioadsorption on activated carbon. Journal of Hazardous Materials. 2009;170:298–305.
- 46. Rahardjo A, Susanto M, Kurniawan A, Indraswati N. Modified ponorogo bentonite for the removal of ampicillin from wastewater. Journal of Hazardous Materials. 2011;190:1001–8.
- Ding R, Zhang P, Seredych M, Bandosz TJ. Removal of antibiotics from water using sewage sludge- and waste oil sludge-derived adsorbents. Water Research. 2012;90:40-6.
- Bui TX, Choi H. Adsorptive removal of selected pharmaceuticals by mesoporous silica SBA-15. Journal of Hazardous Materials. 2009;168:602–8.
- Chang PH, Li Z, Yu TL, Munkhbayer S, Kuo TH, Hung YC, et al. Sorptive removal of tetracycline from water by palygorskite. Journal of Hazardous Materials. 2009;165:148-155.
- 50. Zazouli MA, Mahvi AH, Dobaradaran S, Barafrashtehpour M, Mahdavi Y, Balarak D. Adsorption of fluoride from aqueous solution by modified *Azolla filiculoides*. Fluoride. 2014;47(4):349-58.
- 51. Balarak D, Jaafari J, Hassani G, Mahdavi Y, Tyagi I, Agarwal S, Gupta VK. The use of low-cost adsorbent (Canola residues) for the adsorption of methylene blue from aqueous solution: Isotherm, kinetic and thermodynamic studies. Colloids and Interface Science Communications. Colloids and Interface Science Communications. 2015;7:16–19.
- 52. Kumar PS, Ramalingam S, Senthamarai C, Niranjanaa M. Adsorption of dye from aqueous solution by cashew nut shell: Studies on equilibrium isotherm, kinetics

and thermodynamics of interactions. Desalination. 2010;261:52–60.

- 53. Mane VS, Mall ID. Kinetic and equilibrium isotherm studies for the adsorptive removal of Brilliant Green dye from aqueous solution by rice husk ash. J. Environ. Manage. 2007;84:390–400.
- 54. Balarak D, Bazrafshan E, Kord Mostafapour F. Equilibrium, kinetic studies on the adsorption of acid green 3 dye onto *Azolla filiculoides* as adosorbent. American Chemical Science Journal. 2016;11(1):1-10.
- Balarak D, Mahdavi Y, Bazrafshan E, Mahvi AH. Kinetic, isotherms and thermodynamic modeling for adsorption of acid blue 92 from aqueous solution by modified azolla filicoloides. Fresenius Environmental Bulletin. 2016;25(5):1321-1330.
- Diyanati RA, Yousefi Z, Cherati JY, Balarak D. Investigating phenol absorption from aqueous solution by dried azolla. Journal of Mazandaran University of Medical Science. 2013;22(87):13-21.
- 57. Safa Y, Bhatti HN. Adsorptive removal of direct textile dyes by low cost agricultural waste: Application of factorial design analysis. Chem Engin J. 2011;12(167):35-41.

- 58. Zazouli MA, Mahvi AH, Mahdavi Y, Balarak D. Isothermic and kinetic modeling of fluoride removal from water by means of the natural biosorbents sorghum and canola. Fluoride. 2015;48(1):15-22.
- Ali I. The quest for active carbon adsorbent substitutes: Inexpensive adsorbents for toxic metal ions removal from wastewater. Sepn. & Purfn. Rev. 2010;39(3-4):91-171.
- Balarak D, Mahdavi Y, Maleki A, Daraei H, Sadeghi S. Studies on the removal of amoxicillin by single walled carbon nanotubes. British Journal of Pharmaceutical Research. 2016;10(4):1-9.
- Diyanati RA, Yousefi Z, Cherati JY, Balarak D. The ability of azolla and lemna minor biomass for adsorption of phenol from aqueous solutions. J Mazandaran Uni Med Sci. 2013;23(106):17-23.
- 62. Balarak D, Joghataei A. Biosorption of Phenol using dried Rice husk biomass: Kinetic and equilibrium studies. Der Pharma Chemica. 2016;8(6):96-103.
- Balarak D, Joghataei A, Azadi NA, Sadeghi S. Biosorption of acid blue 225 from aqueous solution by *Azolla filiculoides:* Kinetic and equilibrium studies. American Chemical Science Journal. 2016; 12(2):1-10.

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